

SNAP CODE:	100201
	100202
	100203
	100204
	100205

SOURCE ACTIVITY TITLE:	CULTURES WITHOUT FERTILISERS
	<i>Permanent Crops</i>
	<i>Arable Land Crops</i>
	<i>Rice Field</i>
	<i>Market Gardening</i>
	<i>Grassland</i>

NOSE CODE:	110.02.01
	110.02.02
	110.02.03
	110.02.04
	110.02.05

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1 ACTIVITIES INCLUDED

This chapter considers the emissions of ammonia (NH₃), nitrous oxide (N₂O), other oxides of nitrogen (NO_x) and volatile organic compounds (VOCs). Cultures without fertilisers are soils cultivated for crop production and grasslands, for cutting and grazing, which are not given N-fertiliser (e.g. legumes and grass/clover swards). This includes some grass in hill-land, which is grazed by livestock, as well as lowland grass that only supports small numbers of animals and does not require fertiliser-N. Emissions from the crops and their decomposing residues are also considered. Persistent Organic Pollutants are dealt with separately under SNAP Code 100600, Use of Pesticides.

Emissions following animal manure application are considered in SNAP Code 100500, Manure Management.

This chapter is a development of chapter ag100100 (EEA 1996) which dealt with cultures both with and without fertilisers. Cultures with Fertilisers are now considered in chapter 100100. Reference may be made to that chapter for further discussion of some of the topics covered here.

In this chapter 100206 includes 'Set-Aside' Land.

2 CONTRIBUTIONS TO TOTAL EMISSIONS

The major source of NH₃ emissions in Europe is volatilization from livestock excreta. Ammonia may also be emitted from the application of N-fertilisers and from fertilised crops. Emissions from unfertilised crops are usually considered to be negligible, although there may be some emissions from N-rich legumes.

The greatest proportion of N₂O emitted by agriculture is considered to be by soil processes following the application of N fertilisers and animal manures to land. However N₂O emissions may also take place during the breakdown of crop residues and mineralisation of excretal-N deposited during grazing and soil organic matter. Emission of N₂O may also occur place following the deposition of other N compounds (NH₃ and NO_x) to unfertilised soils.

Soils and crops are regarded as a net sink for most NO_x compounds. However NO may be released from soils during the mineralisation of N from incorporated crop residues and soil organic matter followed by nitrification. Only NO emissions are therefore discussed. At present estimates of the proportion of these emissions that arise from cultures without fertilisers are extremely uncertain.

Table 2.1: % Contribution of total emissions of the CORINAIR94 Inventory (28 Countries) from cultures without fertilisers.

SNAP code	SO ₂	NO _x	NMVOC	CH ₄	CO	CO ₂	N ₂ O	NH ₃
100200	-	0.2	0	1.4	-	-	1.5	0.3

0 = emissions are reported, but the exact value is below the rounding limit (0.1 %).

- = no emissions are reported

As can be seen from Table 2.1, emissions of NH₃, NO and VOCs are all < 1% of current total emission estimates. These do not therefore require a methodology for calculation. However given current uncertainties over the magnitude of emissions from unfertilised crops and grass, some information is given in this chapter. This summarises current understanding and uncertainties.

3 GENERAL

3.1 Description

3.1.1 Ammonia

The direct emissions of NH₃ that have been measured from crops have been attributed to enrichment of the apoplast with NH₄⁺ following addition of fertiliser-N (Sutton et al. 1995). There is very little information on ammonia emissions from cultures without fertilisers. Significant emissions are not expected from unfertilised crops (with the exception of legumes).

Crops of agricultural legumes, while not given fertiliser-N, have been estimated to fix amounts of N as great, or greater than applied as fertiliser to agricultural crops (Sylvester-Bradley 1993). Thus emissions of NH_3 may be expected to be similar to those from fertilised agricultural crops (e.g. $0\text{--}15 \text{ kg ha}^{-1}\text{yr}^{-1}$, Sutton et al. 1995). Data on NH_3 fluxes over legume crops are sparse. Dabney and Bouldin (1985) measured a small net emission of *c.* $2 \text{ kg ha}^{-1}\text{yr}^{-1}$ $\text{NH}_3\text{-N}$ from an alfalfa crop. Harper et al. (1989) found net depositions of $0.4\text{--}3.1 \text{ kg ha}^{-1}\text{yr}^{-1}$ from soybeans. Lemon and van Houtte (1980) measured both emission and deposition fluxes over soybeans.

Some recent results (R. Harrison, ADAS Boxworth, pers. comm.) also suggested no net emission over the growing season. However in that study small ($1\text{--}2 \text{ kg N ha}^{-1}$) emissions early in the season were balanced by deposition ($2\text{--}3 \text{ kg N ha}^{-1}$) later in the season. This deposition may have been a consequence of grazing activity in the locality. The possibility remains that agricultural legume crops, in predominantly arable areas, may emit small amounts of NH_3 . Ammonia fluxes are also bi-directional over fertilised arable crops. These few data suggest that, at present, only an approximate, indicative emission factor for cultivated legumes can be made.

Measurements of NH_3 fluxes over unfertilised grassland have usually shown net deposition of NH_3 (Sutton et al. 1993). Whitehead and Lockyer (1989) measured emissions only from grass foliage with a high-N content where large amounts of fertiliser-N had been applied.

Ammonia emissions from unfertilised grass, grazed by livestock, have been made by Jarvis et al. (1989 and 1991) and Ledgard et al. (1996). Jarvis et al. (1989) found annual NH_3 emissions of $7 \text{ kg ha}^{-1} \text{ N}$ from a grass/clover pasture grazed by beef cattle. This was *c.* 4% of the estimated N fixation by the clover ($160 \text{ kg N ha}^{-1} \text{ yr}^{-1}$), and *c.* 70% of NH_3 emissions from grazed grassland given $210 \text{ kg ha}^{-1} \text{ N yr}^{-1}$. Jarvis et al. (1991) measured NH_3 emissions from pastures grazed by sheep, including an unfertilised clover monoculture. Emissions of NH_3 from the unfertilised grass/clover pasture ($2 \text{ kg N ha}^{-1} \text{ yr}^{-1}$) were less than from an unfertilised grass field ($4 \text{ kg ha}^{-1} \text{ yr}^{-1}$), whilst emissions from the pure clover pasture ($11 \text{ kg N ha}^{-1} \text{ yr}^{-1}$) were greater than from grassland given $420 \text{ kg N ha}^{-1} \text{ yr}^{-1}$. These losses were smaller (by a factor of 3) than from pastures grazed by cattle (Jarvis et al. 1989). Ledgard et al. (1996) measured an annual NH_3 emission of 15 kg/ha from unfertilised grass/clover grazed by dairy cattle. There are considerable uncertainties in generalizing from these limited data. Differences in emission are likely to be the result of variation in temperature, soil type and livestock type. In addition, if unfertilised grassland is cut and left in the field for an extended period, decomposition may result in some emission.

3.1.2 Nitrous Oxide

The methodology adopted by the IPCC may be used to calculate emissions of N_2O from cultures without fertilisers as the sum of, i) direct soil emissions and, ii) indirect emissions. Direct soil emissions from cultures without fertiliser may be the result of biological N fixation, excreta of grazing animals, crop residue incorporation and soil cultivation. Indirect emissions may arise as a consequence of atmospheric deposition of NH_3 and NO_x to unfertilised soils.

In soil N₂O is produced predominantly by nitrification (the oxidation of ammonium (NH₄⁺) to nitrate (NO₃⁻), and denitrification (the reduction of NO₃⁻ to gaseous forms of N : N₂O and N₂). The rate of N₂O production is primarily dependent on the availability of mineral N in the soil (Bouwman 1996). The cultivation of soils, grazing by livestock and incorporation of crop residues are likely to increase soil mineral N concentrations and hence N₂O emission (e.g. Flessa and Beese 1995). Moreover large emissions of N₂O may take place following the thawing of frozen soils (Kaiser et al. 1997).

The magnitude of direct N₂O emissions may be expected to vary with a range of soil and environmental factors. More work is needed on partitioning of N₂O production between nitrification and denitrification. Incorporation of N-rich (e.g. leguminous) residues into moisture-retentive soils produces greater N₂O emissions than from free-draining soils (Skiba et al. 1992). Incorporation into warm soils is also likely to lead to greater emissions than from soils which are cold. Rapid crop growth, and demand for NO₃-N, may be expected to reduce N₂O emissions by reducing the pool of mineral N available for denitrification. Such soil and environmental factors are also likely to influence the magnitude of indirect N₂O emissions following atmospheric deposition of NH₃ and NO_x.

3.1.3 Nitric Oxide

Nitric oxide (NO) may be emitted from unfertilised soils as a consequence of nitrification or denitrification. If soils are maintained above pH 5.0, NO emission is likely to be mainly from nitrification (Remde & Conrad 1991; Skiba et al.1997). Increased nitrification is likely to occur following soil cultivation and incorporation of crop residues. Activities such as tillage and incorporation were considered to increase NO emissions by a factor of 4 (Skiba et al. 1997), for periods of between 1 and 3 weeks.

The main determinant of NO production in agricultural soils is mineral N concentration (Skiba et al 1997); which in unfertilised cultures is increased by residue incorporation and cultivation. As a first approximation, 0.3% of N inputs may be expected to be lost as NO, as per SNAP Code 100100, Cultures with Fertilisers. Thus a knowledge of the available N concentration, and mineralisation rate of crop residues, could provide an estimate of soil NO emissions following cultivation.

However, very little data are available on emissions of NO from unfertilised soils that may be used as a basis for compiling an inventory.

3.1.4 Volatile Organic Compounds

Volatile Organic Compounds (VOCs) are defined as "...all those organic compounds, other than methane, which can produce photochemical oxidants by reaction with nitrogen oxides in the presence of sunlight".

Three categories of sources may be distinguished:

- (1) Activities that emit VOCs by combustion or evaporation;
- (2) Land clearing, including burning;
- (3) Biogenic processes.

The primary sources in the agricultural sector are:

- (1) Burning stubble and other plant wastes;
- (2) The use of organic solvents in pesticide production;
- (3) Anaerobic degradation of livestock feed and animal excreta.

These 3 major sources are dealt with elsewhere. Stubble burning in SNAP Code 100300, emissions from burning other crop residues in SNAP Code 090700 (Open Burning of Agricultural Wastes). Emissions from livestock manures are included in SNAP Code 100500, Manure Management.

The emission of some VOCs may be of benefit to plants to attract pollinating insects. While others may be involved in interactions, be waste products or a means of losing surplus energy (Hewitt and Street 1992). These emissions have been observed to increase when plants are under stress. Factors that can influence the emission of VOCs include temperature and light intensity, plant growth stage, water stress, air pollution and senescence (Hewitt and Street 1992). Emissions of VOCs from plants have usually been associated with woodlands (König et al. 1996). Hewitt and Street (1992) took qualitative measurements of the major grass and crop species in the UK (except for barley, *Hordeum vulgare*). None of the grass species were found to emit isoprene or terpenes. The only crop species producing any significant emissions was blackcurrant (*Ribes nigrum*). However, these workers warned against classifying plants as 'non-emitters' on the basis of limited measurements, as plant growth stage had been shown to be an important factor in emission.

Hewitt and Street (1992) concluded that only c. 700 plant species, mainly from N. America, had been investigated as isoprene or monoterpene emitters. Few of these were agricultural crops, and quantitative data was available for only a few species. Many measurements had been made at temperatures higher than those prevailing in N and W Europe.

3.2 Definitions

Animal Manures. Animal excreta deposited in houses and on yards, collected, either with bedding or without, to be applied to land.

Livestock excreta. Animal excreta deposited at any time, including while grazing.

Unfertilised agricultural grassland. Grassland, to be used for either cutting grass for conservation, grazing or both, to which synthetic N-fertilisers have **not** been applied. Phosphorus or potassium fertilisers may be used.

Crop residues. The unharvested parts of crops that are left on the field and ultimately incorporated into the soil.

Hill Land. grassland in the hills or uplands that is used for grazing agricultural livestock.

3.3 Controls

3.3.1 Ammonia

No measures have so far been proposed to reduce NH_3 emissions from cultures without fertilisers. There are some possible suggestions. The area of legumes could be reduced. However the consequence may be an increase in the area of crops requiring fertiliser-N. Ammonia emissions from these may not be less than from legumes. Pure clover pastures may be replaced by mixed grass/clover. This is unlikely to be of much practical significance, as pure clover pastures are uncommon.

3.3.2 Nitrous Oxide

Nitrous oxide emissions may arise following the incorporation of N-rich crop residues into warm moist soil. A control technique may, therefore, be to avoid incorporating residues in late summer/early autumn and delaying incorporation until late autumn where succeeding crops are to be sown before the onset of winter. This will also have the advantage of reducing the potential for NO_3^- leaching. However leaving N-rich crop residues (e.g. from legumes) on the soil surface will probably give rise to NH_3 emissions as they senesce.

3.3.3 Nitric Oxide

In view of the limited information on the loss of NO from unfertilised soils, no specific control measures are proposed at this stage. However, any measure that reduces mineral N production and input to the soil, as discussed in section 3.3.2, will also reduce loss of NO.

3.3.4 Volatile Organic Compounds

To reduce emissions of VOCs, crop residues should be removed from the field (to be used for animal feed and bedding) rather than be disposed of by burning.

4 SIMPLER METHODOLOGY

4.1 Ammonia

Since legumes are the only arable crops regarded as sources of NH_3 in cultures without fertilisers, a simple estimate of NH_3 emissions may be made by multiplying the known area of legumes with an estimated emission factor of $1 \text{ kg ha}^{-1} \text{ yr}^{-1}$.

The following emission factors ($\text{kg NH}_3\text{-N ha}^{-1} \text{ yr}^{-1}$) are proposed for all unfertilised pastures grazed by cattle, and for lowland pastures grazed by sheep :

Grass/clover,	Cattle 7, Sheep 2;
Unfertilised grass,	Cattle 4, Sheep 4.

These emission factors are taken from studies of grazing emissions by Jarvis et al. (1989 and 1991). Greater emission factors from unfertilised grass than from grass/clover swards, may appear contrary to expectations. Jarvis et al. (1991) were unable to fully explain this

observation, but suggested the different crop canopy structure of grass/clover pastures might reduce NH₃ losses.

The use of the above factors also gives greater apparent NH₃ emissions than are estimated by additions of fertiliser-N of up to c. 200kg N ha⁻¹ in Chapter 100100, Cultures with fertilisers. These anomalies emphasise the lack of data available on NH₃ emissions from unfertilised, grazed grass, and hence the uncertainty of this component of the Emission Inventory.

Hill-land grass grazed by sheep, is not regarded as a net source of NH₃ emission over the year (e.g. Sutton et al.1993).

4.2 Nitrous Oxide

Following the IPCC methodology (IPCC/OECD 1997), N₂O emissions from unfertilised agricultural soils may be calculated as the sum of:

- i. direct soil emissions (1.25% of N inputs are emitted as N₂O-N); (where N inputs are from biological N fixation and crop residues). See IPCC Worksheet 4-5, sheet 1;
- ii. direct N₂O emissions from cultivation of histosols (IPCC Worksheet 4-5, sheet 2);
- iii. direct soil emissions (2% of N inputs) from grazing animals (IPCC Worksheet 4-5, sheet 3);
- iv. indirect emissions following deposition of NH₃ and NO_x (1% of N is subsequently re-emitted as N₂O), or leaching and run-off (2.5% of N leached or run-off, IPCC Worksheet 4-5, sheets 4 and 5).

These input data can be estimated from FAO data (see IPCC/OECD 1997) (Table 4.1).

The default emission factors for the above are given in Table 4.2. More detail may be obtained from IPCC Worksheet 4-5, sheets 1-5.

Prior to estimation of direct N₂O emissions, excretal N returns are reduced by 20% to allow for N lost as NH₃. Information on N excretion by livestock is given in SNAP Code 100500, Manure Management. However those values, may be an overestimate for unfertilised grass as they are averages across a range of production systems and both fertilised and unfertilised grassland.

The N₂O emission may be calculated as :

$$FN_{2O-N} = 0.0125 * \text{'net' N inputs} + 0.02 * \text{'net' N inputs from grazing} + 0.01 * (\text{NH}_3 + \text{NO}_x\text{-N emissions to atmosphere}) + 0.025 * \text{N (leached and/or run off)} \quad (1)$$

where units are generally expressed as kg N ha⁻¹ yr⁻¹.

In the case of organic soils (histosols) an additional emission of 5 kg N ha⁻¹ yr⁻¹ is added (Table 4.2).

Table 4.1: Summary of IPCC source categories (IPCC Guidelines for National Greenhouse Gas Inventories, Volume 2: Workbook, 1997) to be reported as CORINAIR sub-sectors for agriculture.

CORINAIR SUB-SECTOR (snap code)	IPCC N ₂ O SOURCE (IPCC WORKBOOK WORKSHEET)
Cultures with/without fertilisers (100100.100200)	<ul style="list-style-type: none"> - Direct soil emission due to N-inputs excluding manure (4-5, sheet 1, excluding animal waste F_{aw}) - Direct soil emissions due to histosol cultivation (4-5, sheet 2) - Direct soil emissions from grazing animals; pasture, range & paddock (4-5, sheet 3) - Indirect emissions due NH₃ and NO_x emissions from synthetic fertiliser use and grazing animals (4-5, sheet 4, excluding animal waste used as fertiliser) - Indirect emissions due N leaching/runoff from synthetic fertiliser use and grazing animals (4-5, sheet 5, excluding animal waste used as fertiliser)
Manure Management (100500)	<ul style="list-style-type: none"> - Manure management: 6 waste management systems (4-1, sheet 2, excluding pasture, range & paddock) - Direct soil emissions due to manure N-inputs excluding grazing animals (4-5, sheet 1, row for animal waste F_{aw} only) - Indirect emissions due NH₃ and NO_x emissions from animal waste excluding grazing animals (4-5, sheet 4, animal waste used as fertiliser only) - Indirect emissions due N leaching/runoff from animal waste excluding grazing animals (4-5, sheet 5, animal waste used as fertiliser only)

The default values used by IPCC (IPCC/OECD 1997) for the above are given in Table 4.2.

The IPCC guidelines estimate direct soil emissions as a fraction of N input to soils, excluding NH₃ emissions. This ‘net’ N input is calculated for N deposited during grazing as:

$$\text{‘net’ N input (kg N ha}^{-1}\text{)} = \text{Excretal-N input (kg N ha}^{-1}\text{)} * 0.8 \tag{2}$$

To calculate N₂O emissions from N deposited during grazing, the ‘net’ N input is multiplied by the emission factor 0.02 kg N₂O-N per kg ‘net’ excretal N input.

$$\text{FN}_2\text{O (grazing)} = \text{‘net’ grazing N input (kg N ha}^{-1}\text{)} * 0.02 \text{ kg N}_2\text{O-N kg N input}^{-1} \tag{3}$$

Discussion of the suitability of this estimate is given in 4.2 of SNAP Code 100100.

Table 4.2: Default emission factors for N₂O emissions from cultures without fertiliser

SOURCE OF N ₂ O	EMISSION FACTOR
Direct soil emissions	
N inputs (crop residues and biological N fixation).	0.0125 kg N ₂ O-N kg N input ⁻¹
# Excretal-N deposited during grazing.	0.02 kg N ₂ O-N kg 'net' N input ⁻¹
Cultivation of histosols.	5 kg N ₂ O-N ha ⁻¹ .
Indirect emissions	
Emission of NH ₃ and NO _x	0.010kg N ₂ O-N kg ⁻¹ NH ₃ -N and NO _x -N deposited
N Leaching and runoff.	0.025 kg N ₂ O-N kgN ⁻¹ leached or lost by runoff.

Manure N inputs, other than from animals during grazing. are dealt with in SNAP Code 100500, Manure Management.

4.3 Nitric Oxide

An estimate of the amount of crop residues, together with their N concentrations, returned to unfertilised soils, together with an estimate of excretal-N deposited during grazing would provide estimate of NO emissions. Assuming 0.3% of the N returned to the soil is emitted as NO.

5 DETAILED METHODOLOGY

5.1 Ammonia

To provide a more detailed methodology it would be necessary to distinguish between different legume species.

Where data are available on the areas of legumes under cultivation and the extent of typical N fixation by each crop type, national NH₃ emission from this source may be approximately estimated as:

$$\text{Legume emission} = \text{sum all legume species} (0.01 * \text{species N fixation} * \text{area of species}) = (\text{kg N ha}^{-1} \text{ year}^{-1}) \dots \quad (4)$$

Where information on average N fixation rates for different legume species is unavailable for a country, 100 kg N ha⁻¹ year⁻¹ may be used as a first estimate.

Further detail may be provided if estimates are available of NH₃ emissions from crops (e.g. hay), or unfertilised crop residues left on the surface. The effects of different climates on NH₃ emissions both from unfertilised crops, and from their residues, needs to be known. However emissions from unfertilised cultures are likely to be small in relation to emissions from livestock husbandry.

5.2 Nitrous Oxide

No more detailed methodology is proposed for estimating emissions of N₂O. However countries may use their own estimates for any step in the IPCC method if this will increase precision. In particular countries are encouraged to estimate NH₃ losses using the methods described in this chapter, rather than the IPCC default values.

5.3 Nitric Oxide

Consideration of the data available suggest that NO emissions may vary substantially according to the prevailing soil moisture regime. Temperature is also considered to have a significant effect on NO emissions. As there is so little information on soil NO emissions from cultures without fertilisers, it is not appropriate to provide a detailed methodology.

6 RELEVANT ACTIVITY STATISTICS

Information is required on the areas of legumes cultivated and by crop type for the more detailed approach, as well as the area of unfertilised grassland grazed by livestock, and an estimate of N deposited in excreta during grazing.

Information may also be required on the amounts and N concentrations of crop residues returned to the soil. This information may be obtained from national statistics on crop production. The area of organic soils (histosols) under cultivation is also useful. Finally, information is needed on deposition of NH₃ and NO_x to soils.

Where spatially disaggregated inventories of unfertilised culture emissions are required (Section 12), information on the spatial distribution of different legume and other unfertilised crops are required. The distribution of cultivated, but unfertilised organic soils (histosols) will also be needed).

Table 6.1: Summary of activity statistics which may be required for the simple and detailed methodologies.

Activity Statistic		Source
Amount and N concentration of crop residues	By crop type	FAO, See IPCC/OECD (1997)
Amount and N concentration of animal excreta deposited during grazing	By Livestock type	SNAP Code 100500, Manure Management
Area of cultivated histosols		FAO, See IPCC/OECD (1997)
Atmospheric emissions of NH ₃ and NO _x		ETCAE, (1997)
N lost from soils by leaching and runoff		FAO, See IPCC/OECD, (1997)

7 POINT SOURCE CRITERIA

Ammonia, N₂O, NO and VOC emissions from cultures without fertilisers should be treated as area sources.

8 EMISSION FACTORS, QUALITY CODES AND REFERENCES

The emission factors for NH₃ losses from cultures without fertilisers are treated as kg N ha⁻¹ of leguminous crops and grazed, unfertilised grassland. For N₂O, losses are kg N₂O-N kg N⁻¹ returned to the soil in crop residues and excreta deposited during grazing, or N deposited from the atmosphere, or lost by leaching or runoff. Losses of N₂O from cultivated organic soils (histosols) are expressed as kg N ha⁻¹.

Table 8.1: Spreadsheet for calculating nitrous oxide emissions from cultures without fertilisers according to the simple methodology.

	A	B	
N input	N emission kg	N input, kg N yr ⁻¹	N ₂ O emission kg N ₂ O yr ⁻¹ (A * B * 44/28)
Crop residues N	0.0125		
Excretal-N deposited during grazing			
Emission of NH ₃	0.010		
Emission of NO _x	0.010		
N lost by leaching or runoff	0.025		
	kg ha⁻¹	Area (ha)	
Cultivation of histosols	5		

9 SPECIES PROFILES

As with forest NMVOC emissions, biogenic emissions from grasslands consist of a wide variety of species, including isoprene, monoterpenes (alpha-pinene, beta-pinene, limonene, etc), and 'other' VOC. The 'other' VOC (OVOC) species consist of a large number of oxygenated compounds (alcohols, aldehydes, etc.), and have proven difficult to quantify in atmospheric samples. Progress in quantification of OVOC from European vegetation has been made recently (König et al. 1996), although many more measurement data will be required before reliable attempts to inventory specific OVOC can be made.

10 UNCERTAINTY ESTIMATES

10.1 Ammonia

The main uncertainty lies in the magnitude of emission factors for unfertilised grassland and leguminous crops, rather than the areas of unfertilised crops under cultivation, which is probably accurate in most countries to better than ±10%. The overall uncertainty is at least a factor of 5.

10.2 Nitrous Oxide

The processes controlling the emission of N₂O from soils are reasonably well understood but their interactions and hence estimates of emission have not yet been accurately modelled.

The magnitude of crop residues and their N contents are only likely to be accurate to within ± 25%. Wet deposition of N may be estimated to ± 20%, but dry deposition of NH₃ to no more than ± 50% (UKRGIAN 1994). As for NH₃ the main uncertainty lies on the generalisation of emission factors, which are perhaps greater than a factor of 5.

10.3 Nitric Oxide

Much less information is available on factors determining losses of NO from soils. (Available N, temperature and soil moisture are likely to be the main factors). In view of the paucity of data, the overall uncertainty is likely to be greater than a factor of 5.

10.4 Volatile Organic Compounds

Estimates of biogenic VOC emissions for the UK range from 38-211 kt yr⁻¹ total VOCs. Between *c.* 10 and 59 kt yr⁻¹ appear to be of agricultural origin. This compares with the CORINAIR94 estimate of only 2 kt yr⁻¹ for SNAP Code 100100 or < 2% of emission from agriculture and forestry. Thus the emission estimates appear to be uncertain by a factor of 30.

11 WEAKEST ASPECTS/PRIORITY AREAS FOR IMPROVEMENT IN CURRENT METHODOLOGY

11.1 Ammonia

Little data is available on NH₃ emissions from leguminous crops, and it does not allow distinction to be made between species. Measurements of emissions from crop residues after harvest is also lacking. The majority of data on NH₃ emissions from grazed grassland have been made on NW Europe. Emission may be greater in drier and warmer areas, e.g. S. Europe. While more work on the development of mechanistic models, which take into account both physicochemical and biological processes is desirable, the primary interest is in understanding atmospheric budgets rather than in the definition of net emissions. It should also be recognised that there is a very large uncertainty in NH₃ emissions in relation to climate and more work is necessary, in particular in Southern and Eastern European conditions.

11.2 Nitrous Oxide

Current estimates of N₂O emissions are also limited by the use of fixed emission factors. More work needs to be done in the development of process-based models that will allow greater discrimination to be made between soils with different moisture regimes, and between areas of different climate. The localised very high inputs of N and C, from animal excreta, are likely to stimulate N₂O emissions.

Estimates of indirect emissions of N₂O are dependent on accurate estimates of N deposition and N leaching and runoff as long as the uncertainties in these estimates are large, then so too will be estimates of indirect N₂O emissions.

11.3 Nitric Oxide

Very little information is available on NO emissions from any of the aspects discussed in this section. More work on NO emissions from unfertilised grassland, land cultivated with legumes and as a result of crop incorporation is particularly desirable. Localised, very high inputs of N and C from animal excreta, are likely to stimulate NO emissions.

12 SPATIAL DISAGGREGATION CRITERIA FOR AREA SOURCES

12.1 Ammonia

Census data on the location of unfertilised crops and grassland, and particularly the distribution of legume crops may be used.

12.2 Nitrous Oxide

Direct emissions may be spatially disaggregated using census data on the distribution of different unfertilised crops and grassland, together with estimates of the N returned in their residues. Data on the distribution of cultivated, unfertilised organic soils (histosols) may also be included to improve spatial disaggregation. Indirect emissions may also be spatially disaggregated if spatial data is available for N deposition and also for N leaching and run off.

12.3 Nitric Oxide

Emissions may be spatially disaggregated using census data on the distribution of different unfertilised crops and grassland, together with estimates of N returned in their residues.

12.4 Volatile Organic Compounds

In the absence of specific data for VOC emissions from different agricultural crops, there appears to be little scope at present for spatially disaggregating VOC emissions.

13 TEMPORAL DISAGGREGATION CRITERIA

13.1 Ammonia

Almost no information is available to generalise on temporal disaggregation of NH_3 from unfertilised crops. Crop emissions are likely to be greatest during crop senescence and from residues left on the soil surface. Dabney and Bouldin (1985) observed a marked seasonal variation in NH_3 fluxes. Emissions were approximately in balance for most of the year, but emissions were greater in the 10 days after the crop was cut for hay. Harper et al. (1989) noted that absorption of NH_3 took place while the soybean crop was well-supplied with water, while emission of NH_3 took place during drought. Such losses are likely to vary greatly from year to year depending upon environmental conditions. Emissions of NH_3 from grazed grassland will largely take place while animals are grazing, although some emission is likely for a period after the animals have left the field.

13.2 Nitrous Oxide

Some data may also be available on the timing of incorporating crop residues. However, until process-based models have been developed and validated it will not be possible to take account of fluxes of N_2O emission that take place when soil mineral N concentrations, soil water regimes and soil temperature combine to produce favourable conditions for denitrification and N_2O emission by nitrification.

Data will be available, for some countries at least, on the temporal variation in N deposition and N leaching and run off.

As for NH₃, losses may vary greatly from year to year, depending upon weather conditions.

13.3 Nitric Oxide

Losses of NO take place mainly as a consequence of nitrification. Peaks in NO emission are, therefore, likely in the first 1 to 3 weeks following incorporation of crop residues and tillage of soils. Data on all these should be available, for some countries at least. At present, however, there is insufficient data on NO emissions to quantify these effects. Ultimately, as the mechanisms of NO production become better understood, climatic data may also be utilised to assess when soil and weather conditions are favourable for nitrification and hence NO production. In common with NH₃ and N₂O, emissions may vary greatly from year to year, depending upon weather conditions.

13.4 Volatile Organic Compounds

Emissions of VOCs are likely to differ according to crop growth stage and weather conditions. Some temporal disaggregation may be possible, if seasonal variations in emissions by non-agricultural plants can be assumed to be valid for unfertilised crops.

14 ADDITIONAL COMMENTS

Where more detailed methodologies than those described here are used by countries, a detailed description should be given of the methodology used, and comparison made to the results of the methodology described here.

15 SUPPLEMENTARY DOCUMENTS

The main supplementary documentation required for applying the estimates in this chapter are details of spatially disaggregated legume crop and unfertilised grass distributions.

16 VERIFICATION PROCEDURES

There are no direct methods to evaluate total inventory estimates of NH₃ emissions from unfertilised croplands, and verification is dependent on laboratory and micrometeorological field studies of emissions from example situations. In particular, many studies have focused on laboratory measurements and there is a need to provide long-term field measurements using micrometeorological techniques to estimate fluxes over a range of crop types in different climates.

Emissions of N₂O, NO and VOCs cannot be verified except by field studies of emissions from example situations. There is a need to long-term field measurements to estimate fluxes over a range of crop types and climates.

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See in particular: Asman (1992), ECETOC (1994), Holtan-Hartwig and Bøckman (1994), IPPC (1995), Skiba et al (1997) and Sutton et al (1995).

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